

Application of Geo-Information Technology in the Study of Soil and Nutrients Loss by Water Erosion for three Different Land Cover

^{1,2}Mushtak T. Jabbar, ¹C.A.I. Chong-Fa, ¹Shi Zhi-Hua and ¹Wang Tian-Wei
¹Huazhong Agricultural University, Wuhan 430070 (China)

²Department of soil and water, college of agriculture, University of Basrah (Iraq)

Abstract: This research integrated the Revised Universal Soil Loss Equation (RUSLE) and the soil statistical analysis with RS and GIS techniques to quantify soil and nutrient loss risk. The Letianxi watershed, located in the southwest part of Hubei Province of China was taken as a case. A system was established for rating soil erodibility, slope length/gradient, rainfall erosivity and conservation practices. The rating values serve as inputs into a Revised Universal Soil Loss Equation (RUSLE) to calculate the risk for soil degradation processes, namely soil water erosion. Two Landsat TM senses in 1997 and ETM 2002, respectively, were used to produce land use/cover maps of the study area based on the maximum likelihood classification method. These maps were then used to generate the conservation practice factor in the RUSLE. In order to assess the effect of land cover and landscape position on soil nutrient consisting of Soil Organic Matter (SOM), Total N (TN), Total P (TP) and Total K (TK), soil samples were collected from the top to foot of hill slope in the study area. The three categories consisted of a typical land cover structure for each polygon in the study area: Soil under forest cover (T1), soil under agriculture crops cover (T2) and soil under no vegetation cover (T3). Annual loss of total N (8.24 kg ha^{-1}), total P (5.88 kg ha^{-1}) and total K (6.98 kg ha^{-1}), per unit loss of soil (t ha^{-1}), was maximum from the soil under no vegetation cover (T3). The loss of total N ranged between 5.30 and 32.27 kg ha^{-1} , total P ranged between 2.14 and 12.42 kg ha^{-1} , total K ranged from 2.12 to 10.31 kg ha^{-1} whereas organic matter loss varied between 10.65 and $236.16 \text{ kg ha}^{-1}$, from three different land covers. ERmapper and Arc/Info software were used to manage and manipulate thematic data, to process satellite images and tabular data source. Results showed that 110.72 km^2 (27.09%) was exposed to very slight soil loss and 227.01 km^2 (55.55%) was exposed to slight soil loss. This study demonstrates the effectiveness of Geo-information technology in generating a soil and nutrient loss map.

Key words: Soil and Nutrient Loss, RUSLE, RS and GIS Technology

INTRODUCTION

Soil and nutrient loss has been identified as an important factor controlling net primary productivity^[2]. Therefore, characterizing special variability and distribution of soil and nutrient loss in relation to site characteristics including climate, land cover, landscape position and other variables is critical for predicting the rates of ecosystem processes^[3], understanding how ecosystems work^[4] and assessing the effects of future land cover change on nutrients^[1].

Land cover is an Integrator of several environmental attributes which influence nutrients export^[5]. Land cover and soil management practices influence the soil nutrient related soil processes, such as erosion, oxidation, mineralization and leaching, etc.^[6] and consequently modify the processes of transport and re-distribution of nutrients. In non cultivated land uses, the type of vegetative cover is a factor influencing the soil organic matter content^[7]. Moreover, soils, through

land cover change, also produce considerable alterations^[8] and usually diminish soil quality after the cultivation of previously untilled soils^[9]. Thus, land cover and type of vegetation must be taken into account when relating soil nutrient with environmental conditions^[6].

In the sloping highland areas of Letianxi watershed, the soil and nutrient loss is one of the main factors limiting soil fertility and crop yields^[10]. Moreover, the Three Gorges Area of China includes a huge project reported to bring significant benefits to the nation and yet at the same time, necessitates the resettlement of over 1 million people (mostly farmers) to more rugged and isolated areas than their original settlements. It is probably the first time in Chinese history that so many people are relocating to marginal lands, which are located on the slopes, with soil of poor structure and low organic matter content. In general, high soil and nutrient loss rates occur during intense storms. Moreover, intensive cultivation and socioeconomic

Corresponding Author: CAI Chong-Fa, Huazhong Agricultural University, Wuhan 430070 (China)
E-mail: cfcai@mail.hzau.edu.cn

pressure for more land have accelerated the rate of soil erosion on sloping lands^[11]. The objectives of this study are to assessment soil loss by using Geographic Information System and Remote Sensing technology and to determine the effect of land cover on soil nutrient loss include Soil Organic Mater (SOM), Total N (TN), Total P (TP) and Total K (TK) in the Letianxi watershed, located in the Three Gorges Area in Hubei Province of China.

MATERIALS AND METHODS

Materials: The study area, located in the southwest part of Hubei province, lies within longitude E 110° 54'to 111° 08'and from latitude N 30° 50' to 31° 10' and has a total area of 408.62 km² (Fig. 1). Present population density averages 74 persons km⁻². The area is situated between the Three Gorges dam and the Guozhouban dam location in a subtropical zone with a monsoonal climate. The mean annual precipitation is 900--1 500 mm, that occurs between May and September and the annual average temperature is 14--18°C. The landform is gentle in the location area, with the elevation ranging from 497 to 1 800 m. The land surface is mainly yellow red soil derived from granite and sandy shale which are classified into Ultisols, based on the soil Taxonomy of the U.S.D.A, (Soil Survey Staff, 1999). All the five thematic layers are generated in the GIS environment at a scale of 1:50 000. The images were acquired in September, 1997 and in September, 2002. The software used for this study was Arc/Info, Erdas and ERmapper imagine.

Methods: RS and GIS Data Processing: Landsat TM and Landsat ETM images from 1997 and 2002 were used to study the location. In order to perform the geometric correction of the image, several ground control point coordinates had to be collected. The ARC EDIT program was used to collect the location of these points from the layer of irrigation and drainage channels and roads. Erdas software was used for geometric correction of the images. The live-link model between ARC/INFO and Erdas was used to match the image and the thematic map layers. More information was collected or calculated and entered into the GIS system. The ARC/INFO and ArcView system is capable of using different information layers for different purposes. The principle thematic layer is the soil map where, all other information is related to its soil polygons. A master tic file was created, with 30 tic points, for geometric correction. The coordinates were converted to the Universal Transverse Mercator (UTM) system using the ARC/INFO software. ARCEDIT was used to edit each information layer and to assign attributes to each polygon. The Table program was also used to assign additional attributes to soil polygon. Other information layers were transferred from Erdas software to the ARC/INFO system. The JOINITEM function of "TABLES" program was used to have all needed attributes in one Polygon attributes table. A

calculation function was used to compute the Revised Universal Soil Loss Equation (RUSLE) for soil erosion processes. The ARC PLOT program was used to plot maps of soil erosion risk. Erdas software was used to classify and compute the (NDVI). ISO-clustering classification technique was used to classify the (NDVI) image into six vegetation density classes (i.e. 0-10%, 10-20%, 20-40%, 40-60%, 60-80% and 80-100%). The classified image file was converted to ARC/INFO vector format to produce nine polygons (Fig. 2)^[12].

Field Work and Laboratory Studies: Nine polygons were chosen for taking the soil samples for the study of the Letianxi watershed, located in the southwest part of Hubei Province of through the fieldwork (Fig. 1). Four typical randomized samples were taken from the top to foot of hill slope for each polygon by using a manual tool to pluck out the aggregates without agitating it from the soil surface layer and then placed in cans and taken to the laboratory for analysis.

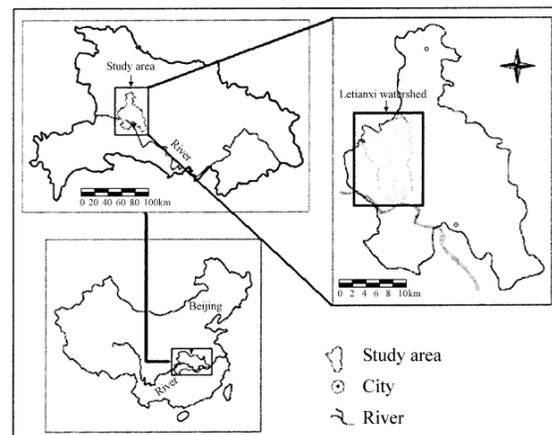


Fig. 1: General Location of Study Area in the (Letianxi Watershed) of South Hubei Province

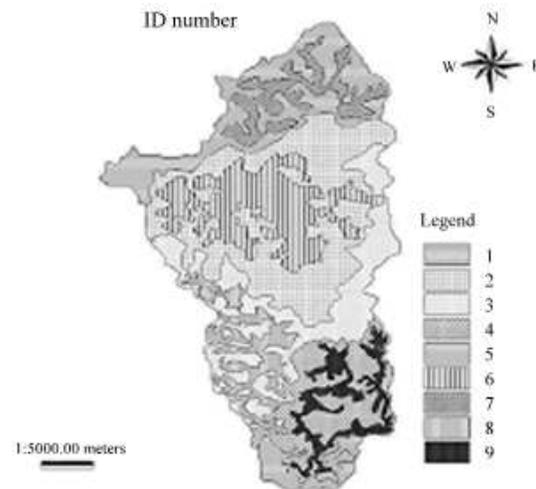


Fig. 2: ID Number in the Study Area

Table 1: Some Chemical and Physical Properties of Soil and their Average Value of the Polygons in Study Area

Physical and Chemical Properties	Polygons (ID number)								
	1	2	3	4	5	6	7	8	9
Soil particle									
Sand (gm Kg ⁻¹)	578.7	670.7	488.9	718.9	734.1	475.2	471.8	307.7	710.9
Silt (gm Kg ⁻¹)	240.4	188.5	340.2	180.8	161.3	344.4	349.3	460.6	188.7
Clay (gm Kg ⁻¹)	180.9	140.8	170.9	100.5	95.6	180.4	178.9	231.7	100.4
Texture	Loamy Sand	Sandy Loam	Loam	Sandy Loam	Sandy Loam	Loam	Loam	Silty Loam	Sandy Loam
Bulk density (g cm ⁻¹)	1.589	1.592	1.385	1.469	1.422	1.382	1.371	1.315	1.425
pH	5.65	5.54	5.45	5.28	5.45	5.86	5.72	5.69	5.91
O.M (gm Kg ⁻¹)	1.92	1.97	0.74	0.47	0.41	0.49	1.91	1.89	0.72
EC (dsm ⁻¹)	0.006	0.005	0.006	0.005	0.005	0.004	0.006	0.004	0.008
Total N%	0.152	0.161	0.048	0.033	0.04	0.035	0.188	0.113	0.056
Total P%	0.028	0.035	0.011	0.006	0.017	0.011	0.069	0.043	0.015
Total K%	1.782	1.617	0.612	2.762	2.181	2.663	1.929	3.118	2.301

Note: 1-Soil particle size: Sand >0.05 mm; Silt 0.05-0.005mm; Clay <0.005 mm

2-Analytical method: Soil texture: hydrometer method; Aggregate size: sieve method

3-Analytical method: O.M: K₂Cr₂O₇+H₂SO₄ → FeSO₄ titration; HCl → NaOH titration

A total 36 sample point was selected based on three categories of land cover. Important physical and chemical properties of the soils were listed in (Table 1) by the procedures of Black *et al.*,^[13]. An extensive field survey was performed throughout the Letianxi watershed using Global Positioning System (GPS) Garmin receiver equipment. The GPS has developed into an efficient GIS data collection technology, that allows for users to compile their own data sets directly from the field as part of 'ground truthing'. The spatial data were digitized from the topographic maps of the study areas by using Arc/view GIS version 3.2a and then the vegetation density attribute data were edited and added to the study databases. The polygons and their attributes were connected with uniform code. Value for average meteorological data were obtained for the investigated location according to information recorded during the period 1997- 2002 from the meteorological station. The present methodology is based upon such parametric models. The Revised Universal Soil Loss Equation (RUSLE) was adopted for the assessment of soil erosion.

Assigning Environmental Factors for Each Soil Polygon: Numerical values were assigned to each soil polygon in the polygon attribute table of the soil coverage layer. The form of the (RUSLE) is:

$$A = F (K * R * LS * C * P) \quad (1)$$

where, A is the soil loss in t ha⁻¹ y⁻¹; F is the function; K is the soil erodibility factor (t. ha. h. ha⁻¹. mJ⁻¹. mm⁻¹); R is the rainfall-runoff erosivity factor in MJ mm ha⁻¹ h⁻¹; L is the slope length factors; S is the slope steepness factor; C is the cover and management factor; and P is the conservation practice factor.

The values of the variable are chosen in such a way that the solving of the equation gives a numerical indication of the degradation rate. This indication is expressed as soil loss in t/h/year for soil water erosion. The formula describes the processes only approximately and the values assigned to each factor are approximate in the present state of knowledge. These values merely give an approximate indication of the magnitude of degradation^[14]. The values were chosen according to the following parameters.

Soil Erodibility Factor (K): The K values were usually estimated using the soil-erodibility nomograph method^[15]. This method uses % clay, % silt (0.002mm – 0.1 mm), % sand (0.1mm- 2mm) and % carbon and soil structure and permeability class to calculate K. However, there is a lack of structure and permeability class data in the soil survey data source. Therefore, we adapted following equation, which is recommended by RUSLE when lacking observation data. The soil erodibility factor was estimated using the equation of William and Renard^[16] which is:

$$K = \{0.2 + 0.3 \exp [- 0.0256 Sd (1 - S i/100)]\} \times [Si / (C 1 + S i)] 0.3 \times \{1.0 - 0.25 C / [C + \exp (3.72 - 2.95 C)]\} \times [1.0 - 0.7 (1 - S d /100)] / \{1 - S d /100 + \exp [- 5.51 + 22.9 (1 - S d /100)]\} \quad (2)$$

where, K, soil erodibility factor (t. ha. h. ha⁻¹. mJ⁻¹. mm⁻¹); Sd, Sand (%); Si, Silt (%); C 1, Clay (%); and C, Carbon (%).

Rainfall-Runoff Erosivity Factor (R): The rainfall-runoff erosivity factor (R) refers to climate (rainfall) factor. The agents for the erosion are raindrops and flowing water. Raindrop, rain splash, or splash erosion

is the process of erosion on barren soil surfaces. The rainfall database obtained from the meteorological office in Letianxi watershed, located in the southwest part of Hubei Province includes the rainfall data for the years 1997 and 2002. Rainfall erosivity was determined by following relationship^[17]:

$$R = \sum_{i=1}^{12} 1.735 \times 10^{1.5 \text{Log} p_i^2 / p - 0.8188} \quad (3)$$

where, P_i is the monthly amount of precipitation and P is the annual precipitation

Topographic Factors (LS): The LS factor was limited of slope $\leq 18\%$ because the data used to develop RUSLE involved slopes up to 18 percent only. However, the study area has 43% of its slope gradient in excess of 30%. Liu *et al.*^[18] employed data from three sites in China with slopes up to 57.7% and reported that the relationship between slope length and soil loss was well approximated by the USLE for computing the L-factor was adopted in this study, i. e. L-factor was described follow:

$$L = (\lambda / 22.13)^m \quad (4)$$

where: m is an exponent that depends on slope steepness, being 0.5 for slope exceeding 5 percent, 0.4 to 4 percent slopes and 0.3 for slopes less than 3%.

To describe the influence of slope steepness, Nearing^[19] produce a single continuous function for S :

$$s = -1.5 + \frac{17}{1 + e^{2.3 - 6.1 \sin \theta}} \quad (5)$$

where θ is the slope angle (degrees).

In order to utilize DEM is calculating LS factor, a program USLE2D.EXE, which is designed to calculate the LS-factor in the RUSLE from a grid-based DEM and provided the user with a number of options in selecting the hydrological flow routing algorithm and the LS algorithm^[20] was used to compute LS factor.

Table 2: C-Factor Value for Different Classes

Land use/cover class	RUSEL-H factor average value
Forest	0.003
Bush and Grasslands	0.01
Weak vegetation	0.60
Sparse vegetation	0.90
Barren soil	1.0
Urban areas	0.0001
Crop lands	0.40
Sand dunes	1.0
Sand lands	1.0
Water bodies	0.0001

Cover and Management Factor (C): The classified land cover map was converted to the C factor layer in RUSLE through reclassification of each land cover type in its corresponding C factor value, which estimated from RUSLE guide tables^[15]. (Table 2) lists the C-factor values for the land use categories. These values were used to re-classify the land cover map to obtain the C-factor for each polygon in the study area. While the RUSLE P-factor reflects the impact of support practices dealing with the average annual erosion rate. It is the ratio of soil loss with contouring and/or strip cropping to that with straight row farming up-and-down the slope. As with the other factors, the P-factor differentiates between cropland and rangeland or permanent pasture. As the study of this research work to estimate soil degradation using RUSLE modeling was applied in the area of non-agriculture or on natural (geological) erosion, it was considered that there was no conservation practice (P) in non-agricultural areas. Therefore as the conservation practice factor (P) value ranges from 0.0-1.0 and the highest value is assigned to areas with no conservation practice, the maximum value for P, which is 1.0, is assigned to this research work area.

Statistical Analysis: Statistical analyses were performed to test the influence of land cover and soil nutrients using one- way ANOVA and mean comparisons were made using the least significant difference (LSD) method with $p < 0.05$. The independent variables used in this study were land cover types and slope aspects. The significance of both of their interactive effect was identified using GLM-MANOVA. All the analyses were conducted through the SPSS program^[21].

RESULTS AND DISCUSSION

Soil Loss and Land Cover Assessment: (Table 3) shows the value of soil loss by water erosion and the input parameters for their calculation. The whole study area is characterized by a present status of high erosivity while the risk is also high. In order to reveal the effect of different conditions, however, recategorization was elaborated. The highest values obtained for the present status and risk are found in the T3 especially in the polygons 8 and 7 (loam to silty loam). All types are subjected to a higher risk of water erosion compared with the present status. The soil is silty loam flats characterized by their Silty texture and poor drainage conditions. Thus, impermeable surface sealing and runoff may occur. These conditions are favorable for both gully erosion and mass movement. Accordingly, values of soil erodibility and soil texture factors are high. The miscellaneous rockland has the characteristics which favor gully formation due to their surface sealing impermeability.

Table 3: Average Value of Soil Loss by Water Erosion and Input Parameters for their Computation for three Different Land Covers for each Polygon

ID	LC	K (t*h/(MJ* mm))	R y MJ* mm/(ha* h)	LS	C and P	Soil Loss (t ha ⁻¹ . yr ⁻¹)	Water Erosion Grade
1	T1 ^{*1}	0.118	423.5	1.077	0.011	0.592	V. slight
	T2	0.125	423.5	1.084	0.011	0.631	V. slight
	T3	0.124	423.5	1.095	0.031	1.788	V. slight
2	T1	0.146	423.5	1.409	0.034	2.962	V. slight
	T2	0.159	423.5	1.458	0.048	4.712	V. slight
	T3	0.177	423.5	1.574	0.087	10.264	Slight
3	T1	0.236	423.5	1.035	0.072	7.447	Slight
	T2	0.238	423.5	1.038	0.072	7.532	Slight
	T3	0.239	423.5	1.038	0.072	7.564	Slight
4	T1	0.141	423.5	1.127	0.037	2.489	V. slight
	T2	0.142	423.5	1.127	0.054	3.659	V. slight
	T3	0.145	423.5	1.127	0.057	3.959	V. slight
5	T1	0.139	423.5	1.132	0.056	3.731	V. slight
	T2	0.145	423.5	1.116	0.056	3.837	V. slight
	T3	0.145	423.5	1.134	0.056	3.899	V. slight
6	T1	0.196	423.5	1.023	0.055	4.670	V. slight
	T2	0.228	423.5	1.033	0.069	6.662	Slight
	T3	0.250	423.5	1.033	0.092	10.061	Slight
7	T1	0.227	423.5	1.540	0.119	17.617	Moderate
	T2	0.231	423.5	1.540	0.113	17.024	Moderate
	T3	0.239	423.5	1.540	0.216	33.669	High
8	T1	0.231	423.5	1.463	0.162	23.185	Moderate
	T2	0.238	423.5	1.471	0.162	24.019	Moderate
	T3	0.252	423.5	1.479	0.162	25.570	Moderate
9	T1	0.164	423.5	1.011	0.055	3.861	V. slight
	T2	0.161	423.5	1.011	0.069	4.704	V. slight
	T3	0.168	423.5	1.011	0.088	6.329	Slight

Numeric range (t ha⁻¹. yr⁻¹): V. slight (0-5), Slight (5-10), Moderate (10-30), High (30-80) and V. high (>80)

*1 (T1) soil nutrients under forest cover, (T2) soil nutrients under Agriculture crops cover and (T3) soil nutrients under no vegetation cover

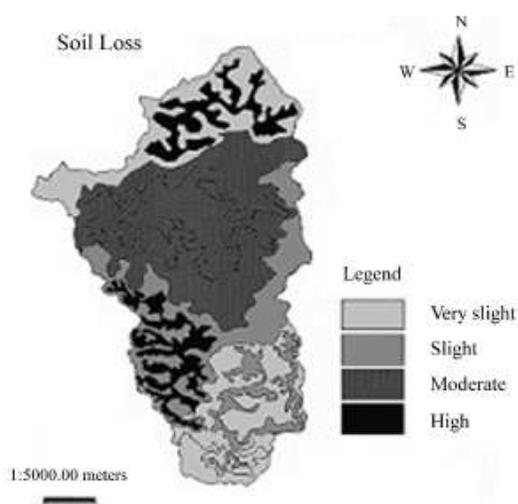


Fig. 3: Risk of Soil Water Erosion in the Location Study

Also, rugged topographic steep slopes and dissected landscape concentrate runoff.

MANOVA results indicated that significant differences in T1, T2 and T3 for soil loss (Table 4). The low value of soil loss in T1 and T2 for each polygon is partly because of the soil under forest and crops cover and their rooting characteristics and partly because of the protection of the soil surface from raindrop impact by the plant canopy^[22]. Forest offers better ground protection than crops. Their cover develops faster than crop area and it is larger to the ground surface cover. The best protection against erosion was attained by intercropping chestnuts (*Castanea mollissima* Bl) and tea (*Camellia sinensis* O. Ktze.).

In (Table 3), we have the general estimation for soil water erosion in the Letianxi watershed, located in the southwest part of Hubei Province of China. It is supposed that all the area is subject of soil erosion as we mentioned previously, mainly by water erosion processes. So we conclude that 82.65% of the study area has very slight and slight soil water erosion, 8.69% has moderate soil water erosion and 8.64% has high soil water erosion (Fig. 3). Without doubt these results show the gravity of the soil water erosion problem with the location study.

Table 4: Soil Nutrient Losses as a Result of Soil Loss

ID	Soil Loss (t ha ⁻¹ .yr ⁻¹)	LC	Eroded Soil			
			SOM (Kg ha ⁻¹)	TN (Kg ha ⁻¹)	TP (Kg ha ⁻¹)	TK (Kg ha ⁻¹)
1	0.592a	T1 ^{*1}	49a	2.42a	1.12a	4.92a
	0.631a	T2	52a	2.59a	1.22a	4.94a
	1.788b	T3	67b	3.81b	1.23a	5.88a
2	2.962a	T1	123a	4.96a	3.29a	9.97a
	4.712a	T2	146a	6.56b	3.47a	12.89b
	10.264b	T3	236b	9.96b	4.77b	13.67b
3	7.447a	T1	158b	4.98a	1.98a	8.77a
	7.532b	T2	161a	5.77a	3.78b	10.12b
	7.564b	T3	163a	9.28b	3.84b	10.96b
4	2.489a	T1	79a	3.17a	2.70a	8.31a
	3.659a	T2	88a	4.22b	2.82a	8.49a
	3.959b	T3	98b	6.19b	4.10b	6.65a
5	3.731a	T1	63a	2.89a	1.69a	8.39a
	3.837a	T2	79b	3.72a	1.73a	8.43b
	3.899a	T3	83b	5.81b	1.87a	8.43b
6	4.670a	T1	127a	2.92a	3.78a	10.89a
	6.662a	T2	219b	4.99a	3.93a	11.68a
	10.061b	T3	227b	9.72b	5.54b	14.69b
7	17.617a	T1	587a	11.23a	7.89a	29.98a
	17.024a	T2	775b	27.99b	12.58b	36.67a
	33.669b	T3	785b	31.85b	14.33b	42.19b
8	23.185a	T1	495a	14.93a	8.99a	20.88a
	24.019a	T3	532a	20.78b	10.43b	31.47b
	25.570b	T3	695b	23.64b	12.89b	34.95b
9	3.861a	T1	108a	6.23a	2.44a	5.96a
	4.704a	T2	119a	6.85a	3.73b	6.28a
	6.329b	T3	149b	8.98b	3.87b	8.84b
F	7.889**		5.657**	4.649**	2.982*	2.691*

The values in each column with the same letter are not significantly (p<0.05, LSD) different among eroded soil

*1 (T1) soil nutrients under forest cover, (T2) soil nutrients under Agriculture crops cover and (T3) soil nutrients under no vegetation cover

Soil Loss-Nutrient Loss Relationship: Soil Organic Matter (SOM), TN and TP were more important than TK (Table 4), the reason being the greater SOM and TN content of soil at study site (Table 1). Both TN and TP are essential for plant growth and hence their loss from runoff and erosion is detrimental to agricultural production, especially where fertilizers are unavailable or too expensive. The study soil contained very little TP because the P was fixed and absorbed in the low pH soil^[23] so, an increase of the soil pH would raise the availability of P for plants. Despite these possible improvements, the reduction of soil erosion probably remains the most effective short-term solution to minimize P losses. Another important aspect of P loss from agricultural land by runoff and soil erosion is the eutrophication of surface water^[23]. In absolute terms, relatively large values of loss of total N are associated with the T3 because of high values of soil loss here. Regression relationships between total N and soil loss are also derived and listed in (Table 4). Relatively higher R2 values obtained for these may be noted. The values of loss of total N per unit loss of soil (i.e. Slope

of the regression line between total N and soil loss) is highest for T3 because of the significant external input of farmyard manure and leguminous crop residue in this land cover system. Inter-comparison of the data on rainfall, soil erosion and nutrient loss in a nature study area having different land covers revealed that of giving soil loss input from the T3 is maximum while soil loss from T1 is the least. As already stated the well terraced topography of the agricultural area is combined with the reasonable crop cover for almost the whole of the year which may cause high interception and hence low surface runoff^[23]. The amount of soil loss per unit amount of runoff is found to be maximum in T3 for all polygons (Table 4). The loss of OM and nutrients (total N; total P and total K) ; per unit loss of soil is found to be maximum in the T3. The aim should be to reduce SOM, N, P and K losses from agricultural fields as effectively as possible using appropriate soil and water conservation techniques. Recent research showed that contour cultivation, grass strips, mulching, terrace cultivation and other methods can effectively reduce soil erosion and hence nutrient loss^[24].

Table 5: Comparisons of Soil Nutrients Change for the three Lands Cover Type

ID	Land covers	SOM (%)	TN (%)	TP (%)	TK (%)
1	T1 ^{*1}	1.95a	0.179a	0.077a	2.028a
	T2	1.74a	0.167a	0.049b	2.012a
	T3	0.97b	0.079b	0.028b	2.010a
2	T1	2.97a	0.295a	0.047a	2.007a
	T2	1.82a	0.244a	0.040a	2.007a
	T3	0.74b	0.051b	0.015b	2.001a
3	T1	1.10a	0.062a	0.023a	2.021a
	T2	0.85b	0.056a	0.022a	2.011a
	T3	0.66b	0.019b	0.012a	2.001a
4	T1	1.37a	0.110a	0.029a	2.021a
	T2	1.00a	0.082b	0.012a	2.010a
	T3	0.44b	0.033b	0.006b	2.001a
5	T1	0.69a	0.087a	0.045a	2.028a
	T2	0.38a	0.080a	0.041a	2.016a
	T3	0.02b	0.020b	0.008b	2.011a
6	T1	0.85a	0.062a	0.017a	2.025a
	T2	0.74a	0.057a	0.017a	2.013a
	T3	0.43b	0.021b	0.016a	2.001a
7	T1	2.33a	0.198a	0.069a	2.001a
	T2	2.07a	0.197a	0.055a	2.010a
	T3	1.59b	0.119b	0.033b	2.010a
8	T1	2.53a	0.540a	0.043a	2.019a
	T2	1.89a	0.413a	0.031a	2.012a
	T3	0.97b	0.125b	0.029a	2.002a
9	T1	0.74a	0.060a	0.053a	2.027a
	T2	0.72a	0.056a	0.051a	2.010a
	T3	0.32b	0.024b	0.026b	2.001a
	F value	4.327**	3.641**	2.852*	2.021

The values in each column with the same letter are not significantly ($p < 0.05$, LSD) different among land covers

*, ** Significant at the 0.05 and 0.01 levels, respectively

^{*1} (T1) soil nutrients under forest cover, (T2) soil nutrients under Agriculture crops cover and (T3) soil nutrients under no vegetation cover

However, the study also showed that the correct choice of agricultural crops offers an effective means to reduce soil and nutrient losses.

Soil Nutrients Losses and Land Cover: The soil nutrients to the location study area differed considerably in SOM, TN, TP and TK (Table 5), the higher SOM and TN contents occurred in soil nutrients under forest cover (T1) and soil nutrients under agriculture crops cover (T2) areas than those under no vegetation cover (T3). Their highest values corresponded to woodland, shrub land and grassland in vegetation cover areas. The results indicate that soil erosion decreases soil nutrient levels, as has been noted by many authors^[25]. There were statistically significant differences in SOM and TN among the three land cover types (Table 5). The mean SOM content varied between 0.32 and 2.97%. Multiple comparisons of SOM

revealed that the SOM level under forest cover and agriculture crops cover was significantly higher compared to under no vegetation cover. Averages of TN (0.024- 0.295%) displayed similar patterns to SOM for multiple comparisons. This similarity may be related to SOM influencing nutrient retention and supply (Brubaker *et al.*, 1993). Mean TP and TK content did not show a marked difference among land covers. High values for TP tended to occur in the orchard and intercropping land with fruit trees. Therefore, differences in soil erosion control of land cover^[26] in such an erosion risky area^[27], may contribute to the significant differences in nutrients for no vegetation cover and vegetation cover areas. In addition, a combination of lower SOM inputs because of less SOM return on land cover, increased aeration by tillage and crop residue collecting partly causes the reduction of TN in cultivated soils^[28]. Unlike SOM and TN there was no significant difference for TP and TK among these land cover. TP content with a narrow range within 0.05 and no differences among land cover may be due to the fact that most of the P is held very firmly in the crystal lattices of largely insoluble forms, such as various CA, Fe and AlPO₄-and also is chemically bonded to the surface of clay minerals^[28]. The high TP content, however, tended to exist in land covers with soil under forest cover. Similar patterns were observed in the three land cover sampling (Table 5).

CONCLUSION

This study assessed the effects of land cover on soil and nutrient loss. Significant differences among land cover were found for most soil nutrients. Soil Organic Matter and TN contents had the higher levels in the forest and crop land than those with no vegetation cover land, while no marked differences in TP and TK occurred among land covers. The soil loss analysis indicated that soil status would suffer from very slight to slight deterioration when the woodland is exploited for agriculture. Under the present land cover management and climate conditions of the study area, the lands with no vegetation cover lands must be abandoned before soil organic matter content is depleted to a the critical value of 0.389% because at that level it is too low to sustain economic yield of crops and erosional processes may be very active resulting in further degradation. The land covers and landscape position resulted in more complex patterns of soil nutrient distribution. Because of the complex nature of soil nutrient patterns largely depending on the land covers and landscape positions, additional research is needed in order to more fully understand the interactive relationships among landscape position, soil erosion, soil nutrient, land cover and its history and management.

ACKNOWLEDGMENT

Natural Science Foundation of China, No.40271073, financially supported this study. The

authors wish to acknowledge Yuling Soil and Water Conservation Bureau for assistance provided in the field survey.

REFERENCE

1. Kosmas, C., S. Gerontidis and M. Marathianou, 2000. The effect of land use change on soils and vegetation over various lithological formations on Lesvos, Catena, Greece, 40: 51-68.
2. Seastedt, T.R., J.M. Briggs and D.J. Gibson, 1991. Controls of nitrogen limitation in Tallgrass prairie. *Oecologia*, 87: 72-79.
3. Schimel, D.S., T.G.F. Kittel, A.K. Knapp, T.R. Seastedt, W.J. Parton and V.B. Brown, 1991. Physiological interactions along resource gradients in Tallgrass prairie. *Ecology*, 72: 672-684.
4. Townsend, A.R., P.M. Vitousek and S.E. Trumbore, 1995. Soil organic matter dynamics along gradients in temperature and land use on the island of Hawaii. *Ecology*, 76: 721-733.
5. Young, W.J., F.M. Marston and J.R. Davis, 1996. Nutrient exports and land use in Australian catchments. *J. Envir. Management*, 47: 165-183.
6. Hontoria, C., J.C. RodrmH Guez-Murillo and A. Saa, 1999. Relationships between soil organic carbon and site characteristics in Peninsular Spain. *Soil Sci. Society of Am. J.*, 63: 614-621.
7. Grigal, D.E. and L.F. Ohmann, 1992. Carbon storage in upland forests of the Lake States. *Soil Science Society of Am. J.*, 56: 935-943.
8. Fu, B., L. Chen, K. Ma, H. Zhou and J. Wang, 2000. The relationships between land use and soil conditions in the hilly area of the loess plateau in northern Shaanxi, Catena, China, 39: 69-78.
9. Jaiyeoba, I.A., 1995. Changes in soil properties related to different land uses in part of the Nigerian semi-arid Savannah. *Soil Land Use and Management J.*, 11: 84-89.
10. Shi, Z.H., 2004. Research on change of landscape pattern and Responses of soil erosion in a watershed level using GIS and RS. Huazhong Agric. Univ.
11. Du, R.H., 1994. The Impact of soil and water losses Upon Ecosystem and Environment in the Three Gorge Area of the Changjiang River. Since press, Beijing.
12. Tucker, C.J., 1979. Red and photographic infrared linear combinations for monitoring vegetation. *Remote Sensing of the Environment*. 8: 127-150.
13. Black, G.R., 1965. Bulk Density. In: Black, C.A., D.D. Evans and J.L. White Eds. *Methods of Soil Analysis*. Part 1. *Agron*, 9: 379-390.
14. United Nations Convention to Combat Desertification (UNCCD), 1994. United Nations Convention to Combat Desertification in Those Countries Experiencing Serious Drought and/or Desertification, Particularly in Africa. New York.
15. Wischmeier, W.H. and D.D. Smith, 1978. Predicting Rainfall Erosion Losses, a Guide to Conservation Planning. Agric. Handbook No. 537, US Department of Agriculture, Washington D.C.
16. Williams, J.R. and E.P.I.C. Renard, 1983. A new method for assessing erosion's effects on soil productivity. *J. Soil and Water Conservation*, 38: 381-383.
17. Ma Jianwen, C.F.M., Y. Xue and Z.G. Wang., 2002. Soil Erosion Monitoring in the Upper Yangtze River Basin of China Using ETM Temporal Data. *IEEE.*, pp: 2365-2367.
18. Liu, B.Y., M.A. Nearing, P.J. Shi and Z.W. Jia, 2000. Slope length effect on soil loss for steep slopes. *Soil Sci. Soc. Am. J.*, 64: 1759-1763.
19. Nearing, M.A., 1997. A Single Continuous Function for Slope Steepness Influence on Soil Loss. *Soil Sci. Soc. of Am. J.*, 61: 917-919.
20. Desmet, P.J. and G. Govers, 1996. Comparison of Routing Algorithms for Digital Elevation Models and their Implications for Predicting Ephemeral Gullies. *Intl. J. of Geographic Information Systems*, 10: 311-331.
21. SPSS Inc., 1993. SPSS for widows base system user's guide release 6.0. Marija J. Norusis/SPSS Inc.
22. Quansah, C., E. Baffoc-Bonnic and F. Agyei, 1990. Runoff and soil loss under four legumes. In ; M.A. Zöbisch (Editor), *Land use and the Environment*. Proc. 11th Annual General Meeting. Soil science Society of Ghana, pp: 67-76.
23. Sharpley, A.N. and S.J. Smith, 1992. Prediction of bioavailable phosphorus loss in agricultural runoff. *J. Environ. Qual.*, 21: 32-37.
24. China National Committee for the Implementation of the United Nations Convention to Combat Desertification (CCICCD), 1996. Country report on combating desertification, The Proc. of the Asia-Africa Forum on Combating Desertification, Beijing. China Forestry Publishing House, pp: 27-52.
25. Lepsch, I.F., J.R.F. Menk and J.B. Oliveira, 1994. Carbon storage and other properties of soils under agriculture and natural vegetation in Sao Paulo State, Brazil. *Soil Use Management*, 10: 34-42.
26. Jiang, Z., Z. Wang and Z. Liu, 1996. Quantitative study of spatial variation of soil erosion in a small watershed in the loess hills region. *J. Soil Erosion and Soil and Water Conservation*, 2: 1-9.
27. Fu, B. and L. Chen, 2000. Agricultural landscape spatial pattern analysis in the semi-arid hill area of the Loess Plateau, China. *J. Arid Environments*, 44: 291-303.
28. Girma, T., 1998. Effect of cultivation on physical and chemical properties of a Vertisol in Middle Awash Valley, Ethiopia. *Communications in Soil Science and Plant Analysis*, 29: 287-295.
29. Brubaker, S.C., A.J. Jones, D.T. Lewis and K. Frank, 1993. Soil properties associated with landscape positions. *Soil Science Society of Am. J.*, 57: 235-239.